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Manuscript:

Unfolding the complexity in water reallocation decision-making in the Heihe River Basin, China

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63 River Basin in Spain) have started to reallocate water back to environment (Dietz et al., 2003; Garrick
64 et al., 2012; Pahl-Wostl, 2015). However, successful cases of water reallocation are relatively sparse,
65 and their expected benefits are rarely met in full (Bouckaert et al., 2021; Marston & Cai, 2016).

66 Water reallocation is a type of collective human decision-making. It refers to changes in historical
67 patterns of water use when the existing allocations are unacceptable hydrologically, economically or
68 environmentally (Garrick et al., 2019; Marston & Cai, 2016). In most cases, water reallocation is a
69 mandatory measure taken by a centralized authority to permanently redistribute existing water
70 entitlements. It is increasingly recognised that water reallocation, on the one hand, has to coordinate the
71 economic development and ecological sustainability in the co-evolved river social-ecological system
72 (Pahl-Wostl et al., 2010); on the other hand, it is deeply constructed by societal values, which is defined
73 as the shared attitudes, cultures, beliefs, ideologies, and routines that construct the collective identities
74 of a society (Nyborg et al., 2016; Roobavannan et al., 2018; Smith & Stirling, 2010). In addition, water
75 reallocation is also influenced by science and technological advances, political structure and
76 institutional arrangements and climate change (Biggs et al., 2015; Eakin et al., 2017). With the
77 increasing connectivity and interactions among these subsystems due to human development, water
78 reallocation decision-making should be a highly complex process.

79 Unfortunately, these interdependencies related to water reallocation have not been taken into full
80 consideration in water reallocation decision-makings, which tended to be linear-thinking and reactive
81 to problematic events (e.g., river dry-up) rather than proactive (Mirchi et al., 2012). Ignoring these
82 interdependencies could hide potential feedbacks underlying each event and create possible time delays
83 which can mask the ultimate effect of management (Meadows, 1974; Sterman, 2001; Wolstenholme,
84 2003), leading to resistance, delay and failure of water reallocation (Gohari et al., 2017; Zare et al.,
85 2019) in many river basins. The ignorance mainly comes from the methodological challenge.
86 Historically the biophysical sciences have been central in the study of environmental problems but at
87 the root of the environmental crisis are human societies and their interaction with the biophysical world.
88 Recently, the importance of including social science in water reallocation, broadly natural resource
89 management decision-making, is increasingly recognised, but as the social sciences often use

90 qualitative approaches, measurement, quantification and hence integration of societal variables
91 including societal values, technology and institutional arrangement into water resources allocation
92 decision making remain a challenge (Wei et al, 2018).

93 This paper aims to develop a system thinking based framework to structurally unfold the complex
94 interactions of water reallocations with the societal, economic and ecological systems in river basins. It
95 will identify weak links, delayed feedbacks and strong external drivers of water reallocation in river
96 basin, thus structurally unfold the complexity in decision-making of water reallocation to the
97 environment. Water reallocation in the Heihe River Basin in China with historical data of over 2000
98 years will be examined, where we have conducted over ten years' cross-disciplinary research, and
99 multiple disciplinary methods and data are accommodated with integrated qualitative and quantitative
100 approaches.

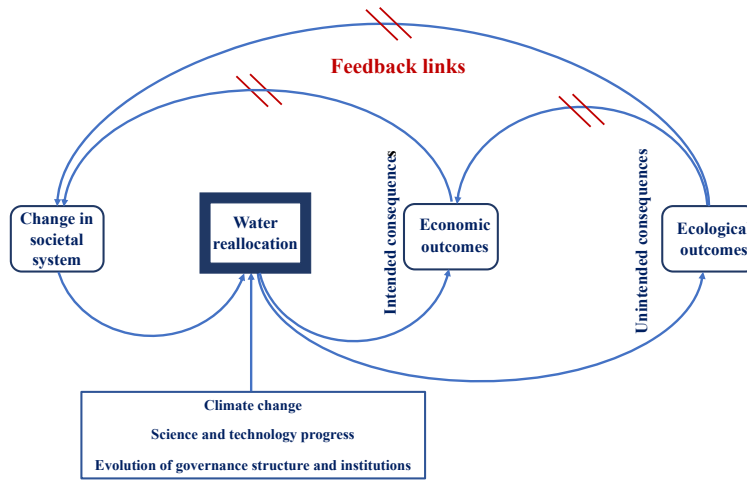
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102 **2. Methods**

103 **2.1 Conceptual framework**

104 We developed a system-based conceptual framework to examine water reallocation decision-making in
105 a river basin (Figure 1). In this framework, a river basin system is defined as a co-evolved social-
106 ecological system according to Ostrom (2009). It consists of the societal subsystem, represented by the
107 values of multiple stakeholders with different objectives and interests which influence water allocation,
108 the economic subsystem represented by economic activities related to water use, and the ecological
109 subsystem represented by the natural ecosystem related to water use. Water reallocation is central to
110 the framework, which is the direct intervention in hydrology that aims to achieve certain economic (in
111 most cases) and/or ecological outcomes and is driven by change in societal values. These three
112 subsystems co-evolve and provide feedbacks to intended and unintended outcomes of water reallocation
113 decisions (Sterman, 2001; Wolstenholme, 2003). The first reciprocal feedback exists between the
114 economic subsystem and the ecological subsystem, which is driven by the changes in share of total
115 available water resources between the two subsystems. The second and third feedbacks exist between

116 the economic and societal subsystems and between the ecological and societal subsystems, respectively,
 117 representing changes in economic and ecological outcomes influence societal values on water
 118 reallocation.



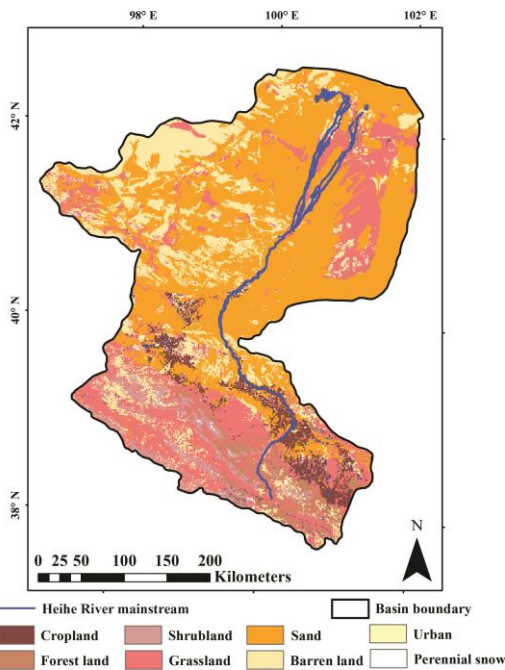
119
 120 **Figure 1. A system developmental framework for water reallocation.**

121 In additions, drivers (internal and external) on water reallocation are considered in this framework.
 122 Climate change influences water reallocation through its direct impact on water availability and
 123 associated social-ecological system. Science and technological progresses influence the science model
 124 we use for understanding water reallocation and our capacity and efficiency on water development and
 125 use. Finally, water acts/regulations are considered as the institutional basis of water reallocation.

126 **2.2 Case study**

127 The Heihe River basin (HRB), located in north-western China, is one of the most arid regions in the
 128 world and an important part of the ancient Silk Road since the Han Dynasty (206 BC–AD 220). The
 129 Heihe River is the second longest endorheic river basin in China, with a length of 821 km and a
 130 catchment area of approximately 130,000 km². The HRB originates from the Qilian Mountains with
 131 annual rainfall of 500 mm and covered with forest and grassland in the northern side of the Tibetan
 132 Plateau, which is the principal water source for the HRB. Its midstream area consists of arid and heavily
 133 cultivated croplands, with mean annual precipitation ranging from 100 to 250 mm. The Heihe River

134 enters its terminal lakes in the downstream, which mainly comprises desert and a small area of natural
135 grassland with a mean annual precipitation of less than 50 mm (Figure 2).



136

137 **Figure 2. The Heihe River Basin** (data source: <https://www.tpd.cn/en/special/heihe/>).

138 The rise and fall of civilizations in the HRB have been closely linked with water. The HRB is an
139 important area for grain production in China and a highly developed irrigation district with an
140 unremittingly agricultural history dating back more than 2000 years ago. Long-term water allocation
141 for agricultural development has significantly modified the catchment vegetation conditions and led to
142 continuing environmental degradation. The implementation of water reallocation to the environment in
143 2000 marked the beginning of re-balancing water use between economic development and ecological
144 sustainability in the HRB but entangled with complex social hindrance.

145 There have been extensive studies on the HRB to support grain security, economic development, and
146 ecological restoration for the national development strategy in different periods. The Eco-hydrological
147 Processes Integration Research in Heihe River Basin, an eight-year (2010-2018) program, was one of

148 the largest experimental catchment science programs in the world (Li et al., 2011). By the end of 2017,
 149 there were over 800 publications on HRB in Web of Science, putting HRB in the top 40 most-researched
 150 river basins globally (Wang et al., 2019). In addition, the HRB has a rich written history of over 2000
 151 years. There are also abundant reconstructed data on climate, hydrology and land use (e.g., Qin et al.,
 152 2010; Xie et al., 2015).

153 2.3 Methods and Data

154 We selected the major indicators to represent the water reallocation, societal, economic, and ecological
 155 subsystems and external drivers in the HRB as defined in Figure 1. They are summarised in Table 1.

156 **Table 1. Indicators for water reallocation and river basin subsystems and their data sources**

Subsystems	Indicator	Data sources
Water reallocation	Water allocation agreements	Zhong et al. (2014) and Ministry of Water Resources, China.
	Human water use	Before 1950s, Lu et al. (2015); After 1950s, Gansu Hydrology Bureau, China and Chinese Bulletin of Rivers and Sediments.
	Environmental water use	Before 1950s, Lu et al. (2015); After 1950s, Gansu Hydrology Bureau, China and Chinese Bulletin of Rivers and Sediments.
Economic subsystem	Population	Before 1950s, Tang et al. (2018) and Lu et al. (2018); After 1950s, Gansu Annual Books, China.
	NPP of cultivated oasis	Extended from Lu et al. (2018).
Ecological subsystem	NPP of natural oasis	Extended from Lu et al. (2018).
	Runoff into terminal lakes	Before 1950s, Lu et al. (2015); After 1950s, Gansu Hydrology Bureau, China and Chinese Bulletin of Rivers and Sediments.
Societal subsystem	Societal value on water	Before 1950s: Zhao (2016), Zhou (2011); and Wang (2014). After 1950s: extended from Xiong et al. (2016)
Drivers (internal and external)	Temperature	Before 1950s, Ge (2011); After 1950s, China Administration of Meteorology.
	Precipitation	Before 1950s, Yang et al. (2014);

	After 1950s, China Administration of Meteorology.
Water science and technology	Before 1950s: Wu (2017); After 1950s: Wang et al. (2019), Gansu Annual Books and Gansu Hydrology Bureau, China, Zhiwang Database, Web of Science Database.
Government structure	Before 1950s: Ding (2019); After 1950s: PRC Government website.
Government acts/regulations	Before 1950s: Ding (2019) and Wang et al. (2017); After 1950s: China State Council departments and China Legislative Affairs Office.

157 Firstly, we qualitatively examined the development history of water reallocation agreements. Then we
158 quantitatively represented water reallocation by the amount of water consumed by the economic
159 subsystem and those kept for the ecological subsystem (environmental water use). These are the core
160 indicators linking water reallocation with the three subsystems and external drivers (Figure 1). For the
161 economic subsystem, population was used as a commonly recognised indicator for human development
162 (Holdren & Ehrlich, 1974), and the Net Primary Production (NPP) of the cultivated oasis in the HRB
163 was also considered as an indicator to reflect human impacts on agricultural-focused terrestrial
164 ecosystems (Haberl et al., 2007). For the ecological subsystem, runoff into terminal lakes was chosen
165 as an indicator to reflect the health of an endorheic river ecosystem, and NPP of the natural oasis in the
166 HRB was also considered.

167 The societal subsystem was characterised with the public attitudes on water reported in historical
168 documents and newspapers. These attitudes reflect the collective societal values in the HRB and were
169 captured via a systematic coding scheme using content analysis. This method was used to code textual
170 data into structured categories including the issues discussed and tones regarding issues. The tones of
171 each document were categorized into two different values: “economic-oriented” or “environmental-
172 oriented” on a normalised scale of 0 (neutral) to 1 (100% economic-oriented or environmental-oriented)
173 to understand their influences on water reallocation. The former focused on issues that addressed the
174 demand of economic and human development such as construction of water storage, and irrigation
175 delivery infrastructure for consumptive use, and the latter about issues such as concerns of river dry-up

176 and vegetation degradation. Detailed descriptions for this method can be found in Supplementary
177 Material A.

178 In additions, the impacts of internal and external drivers on water reallocations were investigated.
179 Temperature and precipitation were included as common indicators to express the impacts of climate
180 change, which work together to modify the hydrological processes and water shared between humans
181 and the environment. The progress of science and technology was represented with the relevant
182 historical documents and recent publications in the Web of Sciences to reflect their impacts on water
183 reallocation. The governance impacts were represented with the position of the water management
184 bureau in the Chinese government structure and acts/regulations on water resources to reflect the
185 institutional basis of water reallocation. Similar to the societal subsystem, the impacts of science and
186 technology and institutional arrangements were coded a scale from 0 (neutral) to 1 (complete economic
187 or environmental perspectives) respectively. Detailed descriptions for the methods on these two
188 variables in Supplementary Material A.

189 The author team, in particular the first author, had conducted nearly ten years' multiple disciplinary
190 studies in the HRB during the implementation period of the Program 'The Eco-hydrological Processes
191 Integration Research in Heihe River Basin'. This paper was built on over twenty journal papers
192 produced from six PhD research projects across climatic hydrology, eco-hydrology, hydrologic
193 economy, societal value on water, government institutions on water and science & technology
194 development on water in the HRB, in which the integrated qualitative and quantitative approaches were
195 adopted. Therefore, a majority of data sources for these selected indicators came from our over twenty
196 journal publications (Table 1). It should be noted that different data sources were used for the majority
197 of indicators before and after 1950s, with the data after 1950s being instrumentally collected in a yearly
198 time scale.

199 We selected eight historical periods for this study based on the data availability and our previous work.
200 It spanned from 0 AD when the first large-scale human settlement occurred in the HRB to 2017,
201 covering over 2000 years (Table 2).

202 **Table 2. Study periods**

Historical period	Time span
Han	0 –220 AD
Wei – Jin Era	221–581 AD
Sui –Tang	582-960 AD
Song (Xi Xia) – Yuan	961-1368 AD
Ming – Qing	1369-1911AD
Republic of China (ROC)	1912-1949 AD
People’s Republic of China (PRC)	1949-2017 AD

203 2.4 Analysis approach

204 We applied system-based thinking to explore the complexity of water reallocation. Events, patterns,
205 structures and drivers characterizing a complex system were analysed in a combined way (Sterman,
206 2001). As complex system behaviours in particular those related to the societal subsystems could not
207 be fully understood only through quantitative analysis (Banson et al., 2016; Forrester, 1992), both
208 qualitative and quantitative approach were used in this study. Firstly, the evolution of each subsystem
209 was described with key events in time as decision makers usually offer solutions based on observed
210 events. Then, the patterns (development phases) of each subsystem were analysed by identifying the
211 abrupt changes in time which provides more meaningful information than events. This was conducted
212 by using a change point detection algorithm if the non-linear relationship better fitted the data pattern
213 {Killick, 2014 #1004}{Samhuri, 2017 #1005}. Following, the systemic structure characterized with
214 the interdependencies between subsystems were analysed with Pearson correlations using the SPSS
215 Statistics 25 software with correlation coefficients to represent the strengths of links between
216 subsystems. Additionally, the drivers of water reallocation were analysed to identify if they were missed
217 from the big-picture perspective by conducting Pearson correlations in a pairwise manner for both
218 internal and external drivers with water reallocation.

219 Finally, our conceptual framework was tested by stepwise linear regression models, of which the
220 societal subsystem was the key determinant of water reallocation, taking into considerations of internal
221 and external drivers. Specifically, in these models, the human and natural water consumptions were
222 dependent variables respectively, and societal value, temperature, precipitation, technology
223 development, and institutions as independent variables. These regression models were assessed by

224 adjusted R² for model fits and t-statistics for significance, and Variance Inflation Factors (VIF) for each
225 variable for potential multicollinearity (Zhang et al., 2007).

226

227 **3. Results**

228 **3.1 Evolution of water reallocation in the HRB**

229 The HRB experienced a long history of water reallocation for human use, beginning with the first large
230 scale human settlement in the Han Dynasty (200 AD) and followed by three water reallocations till
231 2017 (Table 3). The first water reallocation occurred during the Ming Dynasty (1400 AD), which was
232 a response to emerging water conflicts within the economic subsystem in the midstream and aimed to
233 allocate more water to greater cultivated areas. The second water reallocation was to address the
234 intensified water conflicts still within the economic subsystem but between the middle and downstream,
235 by allocating water exclusively to the downstream for 10 days before crop grain harvest every year in
236 the Qing Dynasty (1726 AD).

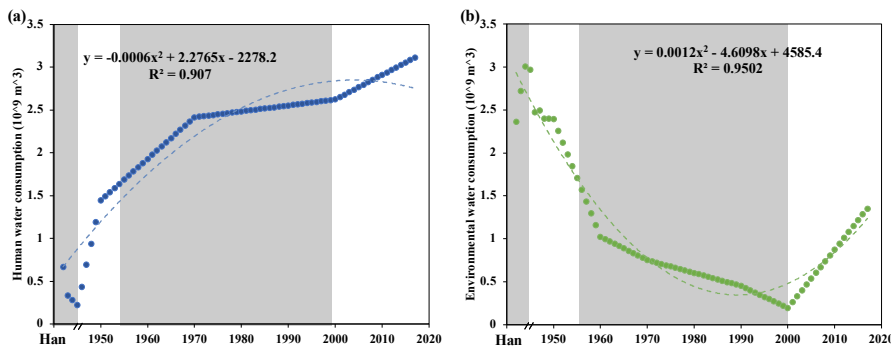
237 The third water reallocation was a response to conflict between human water use and the environment.
238 It was discussed for a long time and experienced great societal hinderance before actual implementation.
239 The initial draft was proposed as early as the 1960s by the downstream community against water
240 overexploitation in the midstream since 1950. With more than 100 dams constructed during this period,
241 there had been river cut-offs in downstream area for over 100 days every year since the 1960s. The west
242 terminal lake Gaxun Nur completely dried up in 1963. While this reallocation draft was under prolonged
243 discussion from 1960 to 1992 involving repetitive bargaining between the midstream and downstream
244 areas, the environmental conditions downstream continued to deteriorate. In the 1990s, river cut-offs
245 extended to over 200 days per year and the east terminal lake Sogo Nur completely dried up in 1992.
246 Although the water reallocation draft was again revised and re-issued by the Ministry of Water
247 Resources in 1992 and 1997 respectively, neither were implemented. It was not finalized and
248 implemented until 2000. This reallocation specified that when the incoming streamflow at Yingluoxia
249 (boundary between the upstream and midstream) was larger than 1.58 billion m³, the runoff into

250 downstream at the Zhengyixia (boundary between the midstream and downstream) should be
 251 maintained at 0.95 billion m³, equivalent to 60% of total water available.

252 **Table 3. Water reallocation in the HRB**

Water allocation	Explanation	Period
Water allocation	The first large scale extraction and use of water by humans.	Han (0-220 AD)
1 st water reallocation	Within economic subsystem in the midstream.	Ming (1400 AD)
2 nd water reallocation	Within economic subsystem between middle and down streams.	Qing (1726 AD)
3 rd water reallocation	Between economic subsystem and ecological subsystem.	2000 AD

253 Regardless of the first and second water reallocations within the economic subsystem during the pre-
 254 industrial period (before 1949), significant increases in human water use in the HRB appeared in the
 255 1950s due to unprecedented industrial development (Figure 3a), and unsurprisingly water for the
 256 environment declined during the same period (Figure 3b). It was not until 1955 that human water uses
 257 surpassed environmental water use at about 1.7 billion m³, marking the first transition in which human
 258 water use began to dominate in the HRB. From the 1960s to 1990s, further deterioration of the natural
 259 environment was signalled by a second phase transition of the amount of environmental water use
 260 reducing to as low as 0.2 billion m³, while the amount of human water use remained stable at about 2.5
 261 billion m³. In the final phase (after 2000), although human water use increased, environmental water
 262 use was also progressively restored to over 1.2 billion m³.



Note: The historical time points before 1950 refer to the average value from the Han to Ming-Qing Period as defined in Table 2, and shadows indicate phase transitions.

263
 264 **Figure 3. Temporal evolution of water reallocation: (a) human water use, and (b) environmental water use.**

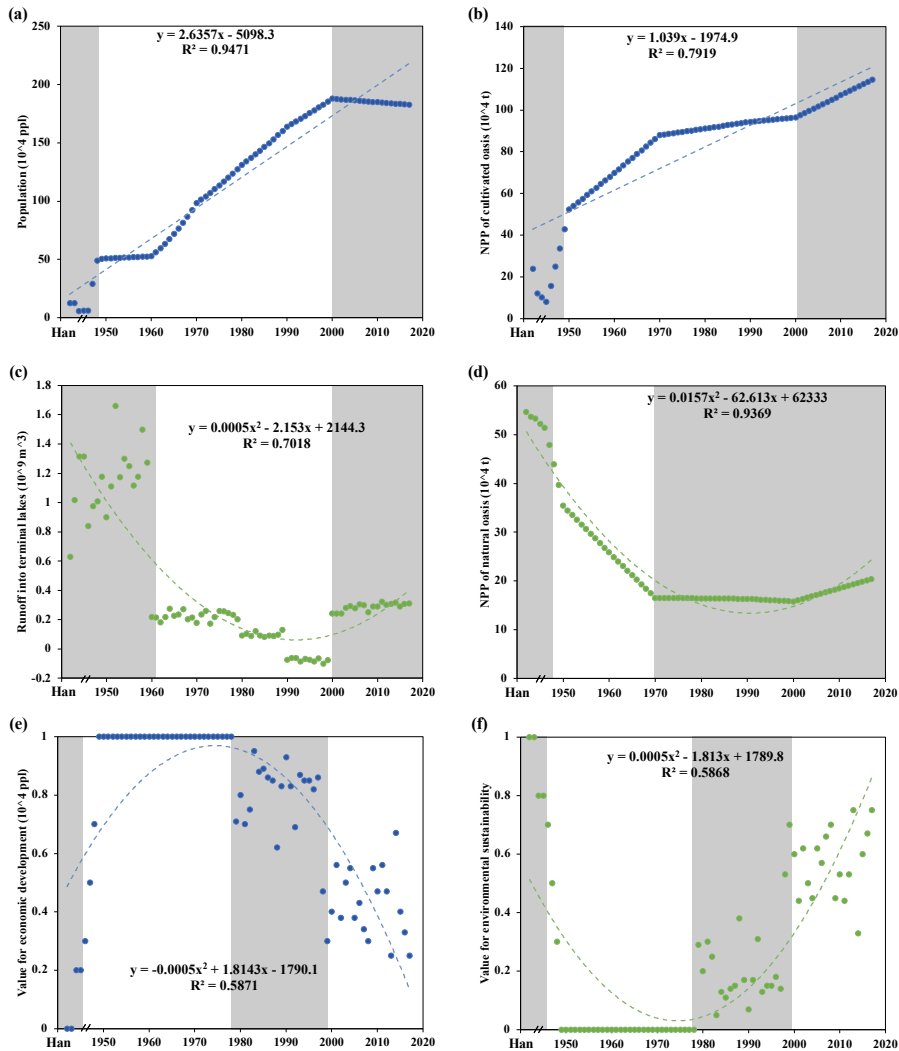
265 **3.2 Evolution of the economic, ecological and societal subsystems in the HRB**

266 The economic subsystem demonstrated clear positive development while the ecological subsystem
267 declined in the HRB (Figure 4a-d). In the economic subsystem, population remained below 200,000
268 during the pre-industrial period, and increased steadily from about 500,000 since the 1950s to about 1.8
269 million in the 2000s (Figure 4a). Similar trend was observed for the NPP of the cultivated oasis, which
270 supported the population growth in the HRB (Figure 4b). With an average value of about 300,000 tons
271 during the pre-industrial period, an increase of about 3 times was observed from the 1950s to 2000s.
272 The final phase was characterized by slight reduction of the population and continued increases in NPP
273 of cultivated oasis.

274 For the ecological subsystem, both runoff to the terminal lakes and the NPP of the natural oasis
275 presented a decreasing trend. The amount of runoffs flowing into terminal lakes varied at an average of
276 about 1 billion m³ per year during the pre-industrial period. A significant phase transition occurred
277 during the late 1950s and early 1960s, when there was over 10-fold reduction in inflow to the
278 downstream areas. The runoffs continued to reduce as the west terminal lake Gaxun Nur dried up in
279 1963 and lake Sogo Nur dried up in 1992. The terminal lakes reappeared after 2000 with reallocated
280 runoffs of about 0.4 billion m³ in 2017, indicating the beginning of a restoration phase (Figure 4c). The
281 NPP of natural oasis dominated the total NPP (up to 90%) during the pre-industrial period and remained
282 relatively stable at about 500,000 tons. A rapid reduction phase occurred in the 1950s with over 60%
283 reductions. Since the 1970s, reductions in NPP of natural oasis slowed down and remained steady at
284 about 200,000 tons (Figure 4d).

285 The evolution of societal values relating to water, characterised in terms of economic development and
286 environmental protection fitted a second-degree parabola (Figure 4e-f). The change of societal values
287 experienced four phases. The first phase spanned across the pre-industrial period to 1911, with societal
288 values inclined towards environmental protection . During this phase, due to extremely low productivity,
289 human survival was subject to the natural environment, humans perceived nature with strong awe. The
290 following phase, extending from the ROC period to late 1970s, was characterised by the transition of
291 an agricultural society into an industrial one. Water resources were considered a mean to support this

292 transition. Environmental sustainability was overshadowed by strong socio-economic goals such as
293 poverty reduction, rural economic development, and national food security. From the 1980s to late
294 1990s, the third phase was characterised by emerging societal concerns regarding the rapidly
295 deteriorating environment. However, the majority of societal value still inclined towards economic
296 development. In the final phase since 2000s, societal values on water waxed and waned between
297 economic development and environmental sustainability. On one hand, large water engineering projects
298 were constructed driven by desires for rapid economic development. On the other hand, frequent water
299 disasters, water pollution and soil erosion remained unresolved and generated strong incentives in
300 environmental management in the society.



Note: The historical time points before 1950 refer to the average value from the Han to Ming-Qing Period as defined in Table 2, and shadows indicate phase transitions.

301
 302 **Figure 4. Temporal evolutions of the economic subsystem: (a) population, (b) NPP of the cultivated oasis;**
 303 **the ecological subsystem: (c) runoffs into terminal lakes, (d) NPP of the natural oasis; and the societal**
 304 **subsystem: (e) value for economic development, (f) value for environmental sustainability.**

305 **3.3 Interdependencies among water reallocation and the three subsystems in the HRB**

306 The interdependences between any two subsystems in the co-evolved river basin socio-ecological
 307 system are summarised in Table 4. Strong correlations were identified between the water reallocation
 308 and economic subsystem, and between the water reallocation and ecological subsystem. Both of
 309 populations (standardised correlation coefficient = 0.89 for 1 unit change of water use) and NPP of
 310 cultivated oasis were significantly ($p < 0.01$) correlated to human water use, while runoff to terminal
 311 lakes (0.84) and NPP of natural oasis (0.95) were significantly ($p < 0.01$) correlated to environmental
 312 water use. These correlations reflect the nature of inland river basins in arid and semi-arid regions, i.e.,
 313 both of the ecological and economic subsystems heavily relied on the amount of surface runoffs
 314 distributed to each subsystem. As a result, there existed strong trade-off relationships between the
 315 economic and ecological subsystems. Population was negatively correlated to runoffs to terminal lakes
 316 (-0.72 , $p < 0.01$), while an even stronger negative correlation was identified between NPP of cultivated
 317 oasis and NPP of natural oasis (-0.94 , $p < 0.01$).

318 It is widely recognised that societal value plays the primary role in the decision on water reallocation.
 319 However, value on water in the HRB did not show statistically significant to human water use (0.07) or
 320 environmental water use (0.2), in another word, societal value on water alone did not play a key role in
 321 water reallocation. Regarding to the feedbacks between the societal subsystem and the economic
 322 subsystem and the societal subsystem and the ecological subsystem, values towards economy was
 323 negatively correlated to population (-0.26 , $p < 0.01$). When population ranged between approximately
 324 500,000 and 1 million, value towards economic development reached 1 (i.e., all values supported
 325 economic development), but then the societal value declined and moved towards environmental
 326 sustainability. Similarly, environmental-inclined societal value was positively correlated to NPP of
 327 natural oasis (0.3, $p < 0.01$), which supported the restoration of natural oasis in the HRB. The effects of
 328 societal value were less responsive to increasing NPP of cultivated oasis (0.058) or restoring runoffs
 329 into terminal lakes (0.003).

330 **Table 4. Interdependencies among water reallocation, economic, ecological, societal subsystems**

Interdependency between subsystems	Correlation coefficient (Pearson's <i>r</i>)
---------------------------------------	---

Water reallocation vs economic subsystem	Human water use vs Population: 0.89 ** Human water use vs Human NPP: 1 **
Water reallocation vs ecological subsystem	Environmental water use vs Runoff to terminal lakes: 0.84 ** Environmental water use vs Natural NPP: 0.95 **
Economic subsystem vs ecological subsystem	Population vs Runoff to terminal lakes: -0.72 ** Human NPP vs Natural NPP: -0.94 **
Societal subsystem vs Water reallocation	Economy inclined value vs Human water use: 0.069 Environment inclined value vs Environmental water use: 0.20
Economic subsystem vs societal subsystem	Population vs Economy inclined value: -0.26 * Human NPP vs Economy inclined value: 0.058
Ecological subsystem vs societal subsystem	Runoff to terminal lakes vs Environment inclined value: 0.003 Natural NPP vs Environment inclined value: 0.30 **
Number of sample n = 76. **: correlation significance at p < 0.01 (2-tailed). *: correlation significance at p < 0.05 (2-tailed).	

331 3.4 Evolution of drivers and their influences on water reallocation in the HRB

332 The climate conditions represented by precipitation and temperature fluctuations remained relatively
333 stable during the entire study period, although there was a tendency of increasing precipitations and
334 temperature since 1950 (refer to Supplementary Material A for details). As a result, temperature did not
335 have statistically significant correlations with either human (0.09) or environmental water use (0.048)
336 (Table 5). On the other hand, precipitations provided water recharge to the HRB and had relatively
337 stronger correlation with human water consumption mainly for irrigation (0.49, p<0.01), whereas
338 negative correlation with environmental water use (-0.34, p<0.01) implying the possibility of an
339 “irrigation efficiency paradox” (Grafton et al., 2018).

340 Technological progresses impacted on water reallocation by developing the capacity of water resources
341 development and improving water use efficiency. The initial phase was characterised by focus on
342 economic development from the Han Dynasty to 1990s, characterised by constructions of large-scale
343 irrigation infrastructures. Science advance was reflected as follows: since the ROC period, hydrology
344 was integrated with engineering for water resources management for economic development. After
345 1960s, hydrology has been widely integrated with economics and engineering in water resources
346 management for economic growth in the most cost-effective approach. With increasing ecological

347 degradations, it was not until about 2000 that more sciences and technologies were developed to
 348 improve environmental sustainability. More water-saving technologies were developed and were
 349 supported by Environmental Science as a major discipline since 1990s (40% of the total publication by
 350 2017) and later supplemented by Meteorology in response to climate change since 2000s. However,
 351 only limited publications were identified in the fields of social sciences (Wang et al., 2019) (refer to
 352 Supplementary Material A for more details). This is evidenced by stronger correlation between
 353 economic development driven technology progress and human water use (0.48, $p < 0.01$), whereas
 354 weaker correlation was identified for environmentally driven technologies on environmental water
 355 consumption (0.2, $p < 0.1$).

356 The central government structure in China remained top-down and the agency/administration for water
 357 resources management remained mostly unchanged in the system for our study period (see
 358 Supplementary Material A for more details). This meant that the impacts of government structure on
 359 water reallocation in the HRB remained unchanged. On the other hand, two phases between economic
 360 and environmental focuses of water acts/regulations were identified. The initial phase spanned from the
 361 Han Dynasty to as late as the 1980s, which was economically driven with a focus on agricultural
 362 irrigation and water development projects. Environmental-protection driven water acts did not appear
 363 until 1980s as the second phase to address environmental problems. However, water acts were still
 364 dominated by more focuses on economic efficiency. This is evident by the stronger correlation between
 365 human water use and economic governance (0.54, $p < 0.01$) than that between environmental water use
 366 and environmental governance (0.31, $p < 0.01$).

367 **Table 5 The influences of external drivers except science model on water reallocation.**

Drivers vs Water reallocation	Correlation coefficient (Pearson's r)
Temperature vs Water reallocation	Temperature vs Human water use: 0.09 Temperature vs Environmental water use: 0.048
Precipitation vs Water reallocation	Precipitation vs Human water use: 0.49 ** Precipitation vs Environmental water use: -0.34 **
Science and technology	Economy inclined technology progress vs Human water use: -0.48** Environment inclined technology progress vs Environmental water use: -0.20

progress vs Water reallocation	
Institutional arrangement vs Water reallocation	Economy inclined water acts/regulations vs Human water use: -0.54 ** Environment inclined water acts/regulations vs Environmental water use: -0.31 **

Number of sample n = 76.

** : correlation significance at $p < 0.01$ (2-tailed).

* : correlation significance at $p < 0.05$ (2-tailed).

368 Finally, linear stepwise regression models on the water reallocation framework indicate that the
369 economic-oriented technological and institutional drivers and societal values were stronger predictors
370 to human water uses ($R^2 = 0.57$) than that for the environmental-oriented drivers to the environmental
371 water use ($R^2 = 0.29$). Specifically, economic-oriented institutions had the strongest influences on water
372 reallocated to human water consumptions (19.57, $p < 0.01$), followed by technology (16.58, $p < 0.01$)
373 and the impacts of economic-oriented societal value were weaker (13.14, $p < 0.01$). On the other hand,
374 environmental-oriented societal values and institutions have similarly strong impacts (16.66 and 16.13,
375 $p < 0.01$), whereas technology didn't have statistically significant impacts on water reallocated to the
376 environment. Additionally, it should note that as the external drivers, climate change represented by
377 temperature and precipitation didn't have statistically significant impacts when taking into
378 considerations of the societal subsystems and the internal drivers on water reallocations in the HRB.

Commented [YW1]: Can you please improve this sentence?

379 **Table 6 Linear stepwise regression models on the water reallocation framework.**

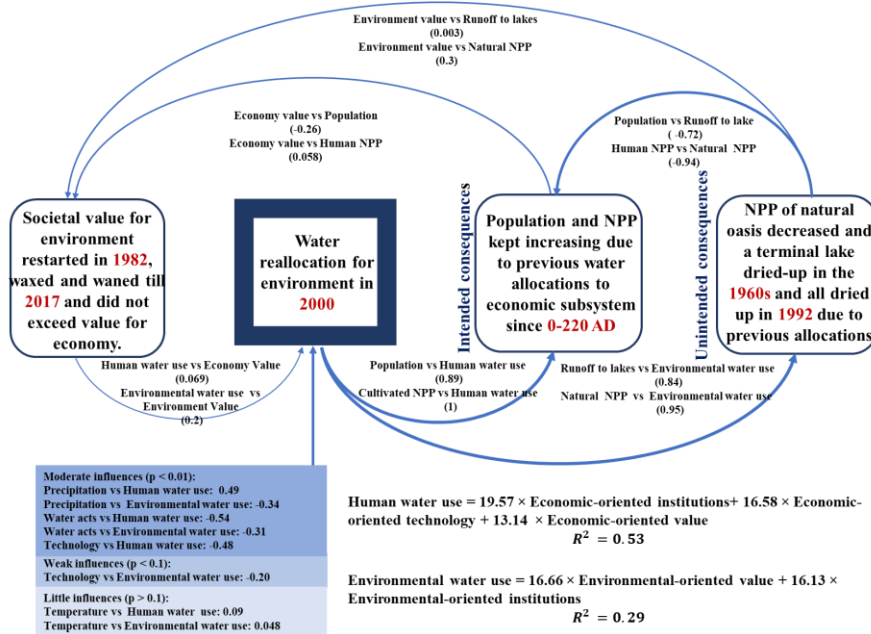
Dependent variable	Coefficients of independent variables (standard error)	t-value	VIF
Human water consumption	Temperature: not significant		
	Precipitation: not significant		
	Economic-oriented institutions: 19.57** (4.31)	4.54	1.87
	Economic-oriented technology: 16.58** (4.38)	5.59	2.11
	Economic-oriented value: 13.14** (2.35)	3.78	1.48
	Overall model fit: $R^2 = 0.53$, $p < 0.001$		
Environmental water consumption	Temperature: not significant		
	Precipitation: not significant		
	Environmental-oriented institutions: 16.13* (4.93)	3.28	1.54
	Environmental-oriented technology: not significant		
	Environmental-oriented value: 16.66** (4.38)	3.81	3.26
	Overall model fit: $R^2 = 0.29$, $p < 0.001$		

Number of sample n = 76.
 **: correlation significance at p < 0.01 (2-tailed).
 *: correlation significance at p < 0.05 (2-tailed).

380

381 4. Discussions and conclusions

382 We developed a conceptual framework for examining water reallocation based on system thinking to
 383 structurally unfold the complexity of water reallocation decision-making, a challenge faced by most
 384 river basins in the world. The Heihe River Basin in China with historical data of over 2000 years was
 385 chosen as the case study. Key findings are summarized in Figure 5:



386

387 **Figure 5. Unfolded complexity in water reallocation decision-making in the HRB**

388 This study found that there was a time lag between ecological responses and economic development.
 389 The first large-scale water allocation for human development in the HRB began during 0-220 AD. With
 390 continuous growth of irrigated agricultural activities and especially constructions of over 100 dams
 391 since the 1950s, ecological degradation characterised by reduced runoff to the terminal lakes and
 392 reduced NPP from the natural oasis was observed in 1960s and worsen since 1990s with zero runoffs

393 into all terminal lakes. This indicates a time lag of over 1000-years between the worst ecological
394 condition reached in 1990s and the initial water resources development in 0 AD, and over 40 years
395 between 1990s and the large scale water resources development in the 1950s. What's worse, societal
396 value that supported environmental sustainability lagged even further behind, only emerging in the
397 1980s which was after two decades since the ecological issues arose and waxed and waned during the
398 most recent decade. Water reallocation was strongly correlated with both the economic and ecological
399 subsystems ($p < 0.01$), and the ecological and economic subsystems were also highly coupled with
400 tradeoff relationships ($p < 0.01$). However, societal value on water in the HRB alone was not statistically
401 correlated with either human or environmental water use. Similarly, neither the economic subsystem
402 nor the ecological subsystem showed strong correlation with the change in the societal values in the
403 societal subsystem. Putting the influences from all drivers and change in societal values together, it is
404 demonstrated that they had more influences on change in the human water use ($R^2 = 0.57$) than that on
405 the environmental water use ($R^2 = 0.29$). Specifically, all of economic-oriented institutions,
406 technologies, and societal values had strong influences on water reallocation to human water uses,
407 whereas only environmental-oriented societal value and institutions were found to be influential on
408 environmental water use.

409 These key findings provide explanations on why the HRB had limited capacity to formulate timely
410 responses of water reallocation to environmental problems. They were: 1) the delayed feedback of the
411 ecological subsystem to development of the economic subsystem; 2) the weak and delayed responses
412 of societal values to ecological degradations and very weak influences on water reallocation to
413 environment; 3) the combined influences of societal subsystem (societal values) and internal drivers
414 (government structures and regulatory institutions, and science and technological development) on
415 water reallocation to the environment is also very limited.

416 The system thinking based conceptual framework developed in this study can assist in investigating
417 interactions, delayed feedbacks and various drivers of the co-evolved socio-ecological system on river
418 basins for a long timeframe. It helps structurally identify the root cause of the problem emerging from
419 quick-fix solutions based on the traditional linear thinking, thus assist formulating strategic-level water

420 reallocation for environmental sustainability (Matthias & Harald, 2010; McDonald et al., 2018).
421 Broadly, it can avoid the environmental crisis caused by human fragmented decision-making. As
422 discussed above, water reallocation incapacity in HRB represents a common challenge facing most
423 river basins in the world, thus, the framework and findings from this study can cast lights onto water
424 reallocation in other river basins.

425 The empirical and quantitative findings from this study are consistent with the recent theoretical and
426 conceptual development for explaining the pervasive social and ecological phenomena in the
427 Anthropocene. These include 1) the inertia (insensitiveness) of societal processes, well noted as
428 “cultural lag” (Rosenschöld et al., 2014); 2) the punctuated equilibrium of ecological systems which is
429 characterized as a long period of stasis being punctuated by more rapid change that disrupts the
430 equilibrium; and 3) “Frenetic standstill” increasingly observed in the societal systems which is
431 characterized by ever accelerating economic changes together with widespread political, cultural and
432 institutional stagnation (the “desynchronization”) (Rosa & Trejo-Mathys, 2013).

433 To conclude, the conceptual framework developed in this study requires accommodation of
434 information/data describing human societal processes, institutions, and development of science and
435 technology paired with those from hydrological and ecosystem processes. Our study has made an
436 empirical step in integrating qualitative and quantitative data from a multi-disciplinary perspective.
437 However, this statistical correlation/regression analysis was only used to reveal the interdependencies
438 among subsystems but would be not used for future prediction. Therefore, continuous efforts on
439 theoretical development are still required to structurally unfold the complexity in decision-making of
440 water reallocation to the environment.

441

442

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446

447 **Author contributions**

448 Y. Wei, A. Western, R. Ison, and M. Sivapalan conceptualised and designed the study, analysed and
449 interpreted the results. S. Wu and Z. Lu collected the data, conducted the data analysis, and interpreted
450 the results. Y. Wei, and S. Wu drafted the initial manuscript, all authors edited and reviewed the
451 manuscript.

452

453 **Declaration of interest**

454 The authors declare no conflict of interest.

455

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